

Porębska G., Ostrowska A. 2018. Nutrient demand and elemental stoichiometry of plants in wetland ecosystems in the Biebrza Valley, Poland. J. Elem., 23(3): 887-899. DOI: 10.5601/jelem.2017.22.4.1496

RECEIVED: 14 July 2017 ACCEPTED: 16 March 2018

ORIGINAL PAPER

NUTRIENT DEMAND AND ELEMENTAL STOICHIOMETRY OF PLANTS IN WETLAND ECOSYSTEMS IN THE BIEBRZA VALLEY, POLAND*

Grażyna Porębska, Apolonia Ostrowska

Institute of Environmental Protection National Research Institute, Warszawa, Poland

Abstract

The study was conducted in the wetland ecosystems in the Biebrza Valley, on peat soils with a varying degree of peat mineralization. The aim of this work was to assess differences between the content of elements in wetland plants and the effect of these differences on the plant elemental stoichiometry. The content of particular elements in a mass unit and the sum of the content of elements in the unit, i.e. in the biological yield, correspond with the nutrient demand of the plants. The content of N, P, K, Ca, Mg, Na and S in mg kg⁻¹ d.w. was determined in aerial parts of various plant species (including leaves of shrubs), and the sum was calculated of all the concentrations of elements in mmol(+) kg⁻¹. The inter-elemental relationships were determined as: (i) an element: element ratio for each element, in (mmol(+) kg⁻¹), and (ii) the percent contribution of each element to the sum of elements. The values of nutrient demand for the species examined varied from about 1000 mmol(+) kg⁻¹ d.w., to more than 6000 mmol(+) kg⁻¹ d.w. Based on significant differences in the values of the sums of element, the species analyzed were split into five groups: up to 2300 mmol(+) kg⁻¹, 2301-3500 mmol(+) kg⁻¹, 3501-4500 mmol(+) kg⁻¹, 4501-6000 mmol(+) kg⁻¹ and above 6000 mmol(+) kg⁻¹, respectively. Meanwhile, the stoichiometry presented in the form of either an element: element ratio or the percent contribution of each element to the sum of elements did not differ significantly between plants with various nutrient demands.

Keywords: peat mineralization, wetland plant species, relationships between elements.

Grażyna Porębska, PhD DSc, Institute of Environmental Protection National Research Institute, Krucza 5/11D, 00-548 Warszawa, Poland, phone: 22 375 05 38

^{*} The material used in this work was acquired under the project HABIT-CHANGE "Adaptive management of climate-induced changes of habitat diversity in protected areas" carried out in the years 2010-2013.

INTRODUCTION

The production of crop biomass in a given space at a given time is driven by multiple factors. All the factors contribute to controlling the rate of biomass production, the quantity of elements utilized for this production, as well as the plant's elemental stoichiometry. According to STERNER and ELSER (2002), the yield understood in the biological sense, as a mass unit produced with the use of a definite quantity of elements, corresponds to the nutrient demand and depends on properties of plant species. According to Liebig's theory of nutrient accumulation in plants, the plant's nutrient demand may be described using the content of individual elements required for producing a mass unit and the sum of the content of all elements in the biological yield (OSTROWSKA 1987).

The physiological role of an element in a plant affects its stoichiometry. At the lowest variable carbon content in plants (THOMAS, MALCZEWSKI 2007, STOLARSKI et al. 2008), it is other elements that determine the relationships between C and the other elements. In both biological yield and crop biomass, the elemental stoichiometry of a plant may be and, in fact, is significantly disturbed, compared to the plant's nutritional requirements, either by the real or supposed deficiencies or excesses of respective elements in the plant's growth environment. The excess of elements is caused mainly by the anthropogenic supply of one element, ignoring the plant's demand for other elements (OSTROWSKA, POREBSKA 2002, 2009).

Nitrogen (N) is a particular element that interfers with the elemental stoichiometry (OSTROWSKA, PORĘBSKA 2009), not only in plants. An inflow of N to the aquatic environment distrubs the balance between N and P, which results in an increased value of the N:P ratio along with an increase in the production of nitrogen-rich toxins by phytoplankton (VAN DE WAAL et al. 2014). According to GÜSEWELL (2005), the negative impact of high N deposition on species diversity in wetlands might be mitigated by P fertilisation.

SARDANS et al. (2012) suggest that plant growth conditions are not the only factor responsible for the relationships between elements, as the latter also depend on properties of a species, its development phase and/or element distribution within a plant, thus on the plant part examined. Plant metabolism was suggested by RIVAS-UBACH et al. (2012) to be the main factor behind the plant's elemental stoichiometry.

Our previous results showed that under the crop's growth conditions the main factor interfering with the plant's elemental stoichiometry is selective fertilization, i.e. using only one or a few elements (OSTROWSKA, POREBSKA 2002, 2009). Under natural conditions, the amounts of nutrients utilized by plants depend on the soil quality and the anthropogenic deposition of various elements. In the case of peat soils, nutrient availability to plants depends mainly on the degree of peat mineralization, as discussed in another paper (SIENKIEWICZ et al., in press).

The aim of this study was to assess differences between plant species of wetland ecosystems, developed on peat soils in the Biebrza Valley, Poland, in terms of nutritional requirements of plants and their effect on the stoichiometry of elements associated with the degree of peat mineralization.

MATERIAL AND METHODS

The study was conducted in the wetland ecosystems of the Biebrza Valley, on peat soils with varying degrees of peat mineralization. The study plots were situated along three transects located in the southern, middle and northern basins of the Biebrza Valley, so as to represent locally typical plant communities and soil properties.

The plant communities varied in terms of the hydrological conditions and degree of peat mineralization in the habitats they occupied. On each study plot, the plant communities were syntaxonomically determined based on the plant species composition.

Soil for analysis was sampled from the three transects at two depths: 0-20 cm and 20-40 cm. The aerial parts of herbaceous plants and leaves of shrubs were sampled from plants growing around the soil profiles. About 100 soil samples, 40 samples of herbaceous plants and circa 10 samples of shrub leaves were analyzed. A summary of the results of soil properties, i.e. the total and dissolved content of macro- and microelements, bulk density, pH, water retention, soil C and N, was previously presented by SIENKIEWICZ et al. (2014). In the present work, we focused on the assessment of plant communities in relation to the peat mineralization, measured by the soil total organic carbon content (TOC).

In the plant samples, after prior preparation (drying, milling), the content of N, P, K, Ca, Mg, Na, S, Mn, Fe, Al, Cu and Zn was determined using the ICP-OES method. In the soil samples, the total content of organic C (TOC) was determined by the combustion method at 1000°C, using a TOC-5000A autoanalyzer (a Shimadzu model), while the total nitrogen content was determined using the Kjeldahl's method.

The content of respective elements in plants is given in mmol(+) kg⁻¹ d.w. The results were statistically analyzed. The variability of the content of each element was analyzed with ANOVA, and significance of differences was determined by the Tukey's test at p = 0.05. The elemental stoichiometry in the biomass was limited to macroelements, with the exception of carbon, and the relationship between the elements was shown as a percentage of each element in the sum of all elements per mass unit. In this way, the inter-elemental relationships could be compared in the biological yield of each species. The relationships in mmol(+):mmol(+) between each pair of elements were calculated in parallel.

Altogether, ca 50 species of plants were analyzed, and 24 species (at least three samples of each) were selected for comparison of differences between species.

RESULTS

The analysis of peat soils in the Biebrza Valley showed significant variations in the degree of peat mineralisation, measured by the TOC content. The content of TOC increased from south to north of the Biebrza Valley, from ca 3% to ca 50%, and this was accompanied by the differentiation of plant communities, from species rich (Transect A) to species poor and even very poor (Transect B) – Figure 1. In all the species examined, the differences in the content of respective elements ranged from 4-5-fold for N, 5-6-fold for P, over 15-fold for K, over 20-fold for Mg, to as much as 40-fold for Ca (Table 1). The coefficients of variation for the mean content of N, P, S and the means of the element sums ranged from 32% to 46%, while those for the content of K, Ca and Mg were from 54% to 88% (Table 1).

The elemental stoichiometry reflects variation in the content of individual elements in the analyzed species, although differences in the values of the element: element ratios are slightly smaller than those in the element content. For example, the N content in the analyzed plants differed 4-5 fold, while the Ca content – approximately 40-fold, whereas the values of the N:Ca ratio varied by 20-30 fold. The values of the element: element ratios, calculated for 53 species on the basis of the average nutrient content, were as follows: N:P – 17.5, N:K – 2.9, N:Ca – 3.1, N:Mg – 5.2, P:K – 0.16, P:Ca – 0.18, P:Mg– 0.30 and Ca:Mg – 1.7.

The mineral composition of the species varied in terms of the content of individual elements, but it was possible to isolate groups of species with similar element content. Based on the values of the sums of the elements in the biological yield, five groups of species were distinguished, which varied in the level of nutrient requirements as follows: to 2300 mmol(+) kg⁻¹, 2301-3500 mmol(+) kg⁻¹, 3501-4500 mmol(+) kg⁻¹, 4501-6000 mmol(+) kg⁻¹ and above 6000 mmol(+) kg⁻¹. Such a division of species was statistically confirmed, since there were significant differences between the above groups in the values of elemental sums, while differences between species inside the groups were insignificant (Table 2).

The percent contribution of the N-content to the sum of elements was found to vary significantly between the above groups of species. Significant differences were also observed in the content of the remaining elements between the plant species which showed the lowest nutrient demand (the sum of elements – up to 2300 mmol(+) kg⁻¹) and between those with the highest demand (the sum of elements > 6000 mmol(+) kg⁻¹) – Table 2.



Fig 1. Vegetation and C content in the layer of 0-20 cm of peat soils in Transect 1 (*a*) and Transect 2 (*b*). Sources: Basic geodata (Topographic map of Poland in the 1:10 000 scale):

 a – plant communities: tall sedge meadows, meadow communities of Molinio – Arrhenatheretea, drying riparian alder-ash wood, drying alder-ash wood Fraxino – Ulmetum in transition to oak-hornbeam Corylo – Piceetum, mixed wood Calamagrostio arundinaceae – Pinetum, alder wood with black currant Ribeso nigri – Alnetum;

b – plant communities: mosaic communities of Magnocaricion and Caricion lasiocarpae, meadow communities of Molinio – Arrhenatheretea

Content of elements in different plant species in mmol(+) $\rm kg^{-1}$

Plant	N	Р	K	Ca	Mg	Na	S	Sum
Aegopodium podagraria	2643	102	764	760	1392	2	131	5794
Angelica sylvestris	2057	138	1046	1148	695	2	312	5398
Asarum europaeum	2071	130	828	1395	420	2	146	4992
Athyrium filix-femina	2077	129	748	306	415	8	169	3852
Calamagrostis arundinacea	1886	113	811	168	205	1	127	3312
Calluna vulgaris	864	33	102	200	134	2	44	1379
Caltha palustris	1571	110	1018	1475	319	n.d.	257	n.d.
Carex acutiformis	1529	76	302	195	186	9	144	2440
Carex nigra	1836	112	323	201	240	2	135	2849
Carex paradoxa (appropinquata)	1186	52	551	247	135	2	103	2276
Carex riparia	1523	94	506	182	190	90	156	2742
Cicuta virosa	2664	118	1603	1395	528	2	178	6488
Comarum palustre	1700	110	556	555	395	11	134	3461
Convallaria majalis	1855	106	805	422	238	2	111	3539
Deschampsia flexuosa	1379	63	263	99	75	1	68	1948
Dryopteris carthusiana	2146	154	553	418	335	8	173	3787
Equisetum fluviatile	1664	177	926	870	262	17	331	4247
Eriophorum vaginatum	1574	93	446	71	134	1	103	2422
Festuca ovina	1207	66	459	70	66	1	82	1952
Filipendula ulmaria	1919	138	617	484	697	4	166	4025
Galeobdolon luteum	2647	141	992	881	582	4	181	5429
Geranium robertianum	2514	134	1108	1065	278	6	191	5296
Geum rivale	1250	100	372	1115	862	2	101	3802
Hepatica nobilis	1736	63	559	710	455	2	150	3675
Humulus lupulus	3586	181	872	1245	653	1	256	6794
Impatiens noli-tangere	2673	152	1195	1384	861	6	190	6461
Lathyrus vernus	2729	87	377	900	281	7	102	4483
Lazula pilosa	1200	84	556	212	265	1	78	2396
Ledum palustre	1093	53	171	309	151	2	55	1834
Majanthemum bifolium	2214	131	856	359	287	2	237	4088
Melampyrum pratense	1871	113	449	535	257	1	124	3350
Menyanthes trifoliata	2436	185	1100	407	215	39	193	4575
Molinia caerulea	1636	53	485	93	136	1	96	2499
Oxalis acetosella	1930	88	603	560	372	3	159	3715
Phalaris arundinacea	1657	50	204	261	85	14	242	2513
Phragmites australis	1989	99	795	130	90	9	197	3309

							cont.	Table 1
Pteridium aquilinum	1364	123	744	157	173	1	89	2651
Pulmonaria obscura	1700	259	1605	665	176	1	130	4536
Ranunculus acris	2014	151	559	985	382	45	178	4315
Ribes nigrum	2116	139	445	1133	361	2	315	4511
Rubus caesius	2471	146	746	388	669	2	152	4575
Rubus idaeus	2193	93	532	924	519	3	113	4377
Rumex obtusifolius	3136	182	1505	456	447	5	305	6036
Salix aurita	1829	120	364	650	257	2	238	3460
Sambucus nigra	3764	137	785	960	873	1	297	6817
Scirpus silvaticus	1693	89	764	210	144	4	168	3072
Stellaria holostea	1969	91	842	525	424	1	146	3998
Thelypteris palustris	1604	81	389	308	491	3	107	2983
Trientalis europaea	1314	85	897	219	368	1	126	3012
Urtica dioica	2082	78	333	3080	770	3	320	6666
Vaccinium myrtillus	1180	61	234	381	177	2	103	2138
Vaccinium uliginosum	1136	50	168	179	136	1	97	1767
Vaccinium vitis-idaea	857	45	137	265	127	1	56	1488
Mean $(n = 53)$	1904	109	660	610	366	7	162	3824
SD	618	43	355	533	262	14	74	1453
V(%)	32	40	54	88	72	219	46	38

SD - standard deviation; V(%) - coefficient of variation

The differentiation between the species with various nutrient demand, assessed according to the content of elements (Table 2), and the percentages of elements in their sum (Table 3), was slightly varied. The percentages of elements in their sum reflect the overall stoichiometry of elements in their mutual interactions. For example, an increase in the share of N in the sum of elements was concurrent with a decline in the share of the remaining elements and vice versa; this regularity is particularly evident in the case of N and Ca.

The comparison of particular species in terms of percentages of elements in their total sum showed similar relationships as did the comparison of groups of species having different nutrient requirements. Significant differences between the shares of the individual elements in their sum were found mainly for the N contribution (Table 4). At the high (> 60%) share of N in the element sum, the share of other elements decreased, particularly the share of Ca (up to several %) and, conversely, when the contribution of N was small, down to approximately 30%, the contribution of Ca increased, even as high as 46 % (Table 4).

Table 2

Sum of elemental content	Number	Ν	Р	К	Ca	Mg	Na	S				
mmol(+) kg ^{.1}	of samples		(mmol(+) kg ⁻¹)									
< 2300	8	1128 a	57 a	250 c	285 a	147 a	1 a	86 a				
2301-3500	16	1641 b	91 b	536 a	199 a	211 a	20 a	133 b				
3501-4500	13	1965 c	113 c	$655 \ a$	$584 \ b$	$428\ b$	6 a	$159 \ b$				
4501-6000	9	$2310 \ d$	141 d	894 b	963 c	517 b	20 a	215 c				
> 6000	6	2800 e	138 d	1021 b	$1575 \ d$	749 c	4 a	243 c				

Mean content of elements in selected groups of plant species in mmol (+) kg⁻¹

Means followed by the same letters do not differ significantly at p = 0.05.

Table 3

Mean share (%) of elen	ments in the sum	of elemental o	content in five	e group of plant	species
------------------------	------------------	----------------	-----------------	------------------	---------

Mean sum total	N	Р	K	Ca	Mg	Na	S
mmol(+) kg ⁻¹				(%)			
1889 a	57.7 c	2.8 b	12.3 a	15.2 bc	7.6 ab	0.1 a	4.3 a
3086 b	58.3 c	3.2 b	18.8 b	6.9 a	7.4 a	$0.7 \ b$	4.7 a
3949 c	50.3 b	2.9 b	16.8 b	14.9 b	10.9 c	0.1 a	4.1 a
4994 d	$45.7 \ a$	2.9 b	17.8 b	19.0 cd	9.9 bc	0.4 ab	4.3 a
6569 e	42.9 a	2.1 a	15.8 b	24.0 d	11.5 c	0.1 a	3.7 a

Means followed by the same letters do not differ significantly at p = 0.05.

Changes in the relationships between N and the remaining elements as well as other pairs of elements were observed. The highest differentiation was in the case of the following ratios: N:Ca, P:Ca and K:Ca (coefficients of variation: 87, 87 and 94%, respectively), while the lowest one was in the case of ratios: N:P and N:S (coefficient of variation: 21 and 27%, respectively) – Table 5.

DISCUSSION

The elemental stoichiometry is governed by plant metabolism, i.e. the physiological role of each element (RIVAS-UBACH et al. 2012), and, in accordance with the ecological stoichiometry theory, the amount of individual elements consumed for the production of biological yield, i.e. mass units (STERNER, ELSER 2002). In other words, these are the nutritional requirements of a plant relative to each element and, according to Liebig, relative also to

4
Ð
6
്ര
H

Sum of elemen	ital content and t	he share of el	ements in the	eir sum in sel	ected plant sp	ecies (mean	values)	
Plant species	Sum	Ν	Р	К	Ca	Mg	Na	\mathbf{s}
	$(mmol(+) kg^{-1})$				%			
Angelica sylvestris	5398	38.1 ab	$2.5 \ bcd$	19.4 ef	$21.4 \ gh$	$12.8 \ efg$	0.03 a	5.8 de
Athyrium filix-femina	3853	$53.8 \ efgh$	$3.4 \ de$	19.4 ef	$8.2 \ abcd$	$10.6 \ def$	0.20 a	$4.4 \ cd$
Calamagrostis arundinacea	3312	$56.9 \ fghi$	3.4 de	24.5 f	5.1 abc	$6.2 \ abc$	0.04 a	3.9 abc
Carex nigra	2850	$64.4 \ ijk$	3.9 e	$11.3 \ abc$	$7.1 \ abcd$	8.4 bcde	0.08 a	$4.8 \ cd$
Carex paradoxa (appropinquata)	2276	$52.1 \ defg$	$2.3 \ abc$	24.2 f	$10.9 \ bcde$	5.9 abc	0.07 a	$4.5 \ cd$
Carex riparia	2742	$55.6 \ fgh$	$3.4 \ de$	$18.5 \ def$	$6.7 \ abcd$	$6.9 \ abc$	$3.32 \ b$	$5.7 \ de$
Convallaria majalis	3540	52.5~defgh	$3.0 \ bcde$	22.7 f	$11.9 \ cde$	$6.7 \ abc$	0.05 a	3.2 ab
Dryopteris carthusiana	3788	$56.8 \ fghi$	$4.1 \ e$	$14.7 \ bcde$	$10.9 \ bcde$	8.8 bcde	$0.21 \ a$	$4.6 \ cd$
Eriophorum vaginatum	2423	65.0 jk	3.9 e	$18.4 \ def$	2.9 a	5.5 ab	0.05 a	$4.2 \ bcd$
Filipendula ulmaria	4025	$47.4 \ cde$	3.4 de	$15.4 \ bcde$	$12.2 \ cde$	17.4	$0.10 \ a$	$4.1 \ abcd$
Galeobdolon luteum	5429	$48.2 \ de$	$2.7 \ bcd$	$19.0 \ def$	16.1 <i>efg</i>	$10.5 \ def$	0.07 a	3.4 ab
Impatiens noli-tangere	6461	$41.4 \ bc$	$2.4 \ bc$	$18.4 \ def$	$21.5 \ gh$	13.4 fg	0.09 a	2.9 ab
Ledum palustre	1835	$59.5 \ hijk$	$3.0 \ bcde$	9.3 ab	$17.2 \ efgh$	$8.1 \ bcd$	$0.09 \ a$	2.9 ab
Molinia caerulea	2499	$65.3 \ k$	$2.1 \ ab$	$19.2 \ def$	$3.9 \ ab$	5.6 ab	$0.04 \ a$	$3.8 \ abc$
Oxalis acetosella	3714	$52.1 \ defgh$	$2.4 \ bc$	$16.1 \ cde$	$15.1 \ ef$	$10,0 \ cde$	$0.08 \ a$	$4.3 \ bcd$
Phragmites australis	3309	$60.1 \ hijk$	$3.0 \ bcde$	24.0 f	$3.9 \ ab$	2.7 a	$0.28 \ a$	$6.0 \ de$
Ribes nigrum	4511	$47.3 \ cd$	$3.1 \ cde$	$10.1 \ ab$	$24.8 \ h$	$7.9 \ bcd$	0.04 a	$6.9 \ e$
Rubus idaeus	4377	49.7 <i>def</i>	2.1 a	$12.1 \ bc$	$21.6 \ gh$	$11.9 \ efg$	0.08 a	2.6 a
Stellaria holostea	3998	$49.3 \ def$	$2.3 \ bc$	$21.0 \ ef$	$13.1 \ de$	$10.7 \ def$	$0.03 \ a$	$3.7 \ abc$
Thelypteris palustris	2983	53.8~defgh	$2.7 \ bcd$	$13.1 \ bcd$	$10.3 \ abcde$	$16.5 \ gh$	0.10 α	3.6 abc
Urtica dioica	6667	$31.2 \ a$	1.2 a	$5.0 \ a$	46.3~i	$11.5 \ defg$	0.04 a	$4.8 \ cd$
Vaccinium myrtillus	2138	$54.4 \ fgh$	$2.9 \ bcd$	10.6 ab	$19.1 \ fgh$	$8.4 \ bcd$	0.07 a	$4.6 \ cd$
Vaccinium uliginosum	1768	64.0~ijk	$2.7 \ bcd$	9.5 ab	$10.3 \ abcde$	$7.8 \ bcd$	0.04 a	$5.6 \ de$
Vaccinium vitis-idaea	1488	$57.5\ ghij$	$3.0 \ bcde$	$9.0 \ ab$	$18.1 \ efgh$	$8.6 \ bcde$	0.06 a	$3.8 \ abc$
		5			•	e,	í	

If next to the mean values at least one letter is the same the species difference between them is not significant (P < 0.05).

895

Table 5

Parameter	N:P	N:K	N:Ca	N:Mg	N:S	P:K	P:Ca	P:Mg	K:Ca	K:Mg	Ca:Mg
Min.	13.9	2.0	0.7	2.7	6.5	0.09	0.03	0.10	0.11	0.43	0.53
Max.	31.1	6.7	22.2	22.2	19.9	0.35	1.31	1.11	6.28	8.87	4.00
Mean	19.3	3.8	6.2	7.0	13.0	0.20	0.33	0.37	1.94	2.17	1.51
V(%)	21	41	87	60	27	40	87	58	94	82	54

Stoichiometry of elements [mmol(+):mmol(+)] in plant species (listed in Table 4, n = 24) with different nutrient demand

V(%) - coefficient of variation

the sum of elements (the so-called Liebig's barrel), consumed for building a mass unit (OSTROWSKA 1987).

The analysis of the aerial parts of ca 50 plant species, from the wetland communities in the Biebrza Valley, showed considerable variation between plant species in terms of their nutrient demand, i.e. the sum of consumed elements, which ranged from below 1000 mmol(+) kg⁻¹ to more than 6000 mmol(+) kg⁻¹. A decreasing trend was observed in the nutrient demand of wetland plant species with the increasing TOC content in peat soils, i.e. a decrease was observed in the quantity of elements accumulated in the biological yield. A similar correlation was noted by WANG and MOORE (2014). The study by LAWNICZAK (2011) showed that N deficiency was the main factor limiting plant growth in wetlands, and similar results were obtained in plants growing on less mineralized soils in the Biebrza Valley.

Based on the statistical analysis, the species were divided into five groups, significantly differing in their nutrient requirements. Species with similar nutrient requirements could hardly be linked with specific plant families. In the previous studies (OSTROWSKA, POREBSKA 2002), significant differences with respect to nutrient requirements were found between the species belonging to such families as: *Poaceae, Chenopodiaceae, Fabaceae* and *Cruciferae*, although there were significant differences between species within the respective families. Generally, the monocotyledonous plants consumed less of elements than the dicotyledonous species to produce a mass unit. For example, GRZEGORCZYK et al. (2013) observed the ability of dicotyledonous plants to accumulate high concentrations of Ca and Mg, which resulted in low K:Ca and K:Mg ratios. Moreover, rapidly growing species consume more elements in a mass unit than slow growing species (POORTER, BERGKOTTE 1992).

We observed that as the TOC content in peat soil decreased, the number of species with ordinary (3000-4000 mmol(+) kg⁻¹) and high (> 4000 mmol(+) kg⁻¹) nutrient demands increased. Peat mineralization results in a decrease of the TOC content and an increase of the nutrients available to plants. It may be suggested that the C content in peat soils seems to be an indicator of the ability of plants to utilize elements. There are more useful indicators (SIENKIEWICZ et al. 2014, SIENKIEWICZ et al. in press) for the assessment of the present peat mineralization or the soluble forms of C and N (DOC and DON). The ratio of C:N, however, suggests only the direction of the mineralization process (OSTROWSKA, POREBSKA 2015).

Differentiation of the elemental stoichiometry in various plant species is small, irrespective of the plants' nutritional requirements (OSTROWSKA, PORĘBSKA 2002). In the present study, slight differences were observed between the species with the smallest and the highest nutrient demand in the relationships between particular elements. This may be the result of these species growing on soils with various content of elements available for plants.

Relationships between elements in plants may be disturbed by various factors, mainly by deficiency or excess of particular elements in the plant growth habitat. The input of nitrogen to the environment (fertilization, deposition) seems to be the main factor that interferes with the relationships between elements. The reasons could be the physiological role of N in the biomass production and excessive input of N with fertilization. According to XIA and WAN (2008), the elevated N input to the environment (fertilization, deposition) in the last century has resulted in an increase of the N content in almost 30% of the analyzed 456 plant species. Notwithstanding the differences in plant species and growth conditions, the results obtained by HAN et al. (2011) on the elemental stoichiometry of 1900 plant species in China do not differ much from the relationships between the elements found in our study.

Numerous authors suggest that the relationships between C, N and P are the most important ones for the functioning of the different groups of organisms. Taking into account little variation of the C content in plants, we disregarded carbon in our study, as the ratios of C to the other elements are controlled by the remaining elements. It seems equally important that, in addition to the C:N:P ratio, the disturbed relationships between different elements in the environment may adversely affect the functioning of producers and consumers. Negative effects may occur in both natural ecosystems (semi-natural) and in crops, although the extent to which they are manifested and appear measurable may vary.

CONCLUSIONS

The results indicate that:

- wild plants growing under natural conditions (or those close to natural) vary in terms of their nutrient requirements, from < 1000 mmol(+) kg⁻¹ to > 6000 mmol(+) kg⁻¹ of the sum of macronutrients accumulated in a mass unit;
- the occurrence of species having different nutrient requirements is associated with the nutrient availability, i.e. the degree of peat mine-ralization;

- the examined plant species with different nutritional requirements do not differ significantly in terms of their elemental stoichiometry; however, there were differences with regard to the inter-elemental relationships between the groups of species with small and high nutrient demands.

ACKNOWLEDGEMENT

We would like to thank PhD Jadwiga Sienkiewicz for phytosociological analyses of floristic material.

REFERENCES

- GRZEGORCZYK S., ALBERSKI J., OLSZEWSKA M. 2013. Accumulation of potassium, calcium and magnesium by selected species of grassland legumes and herbs. J. Elem., 18(1): 69-78. DOI: 10.5601/ jelem.2013.18.1.05
- GÜSEWELL S. 2005. High nitrogen : phosphorus ratios reduce nutrient retention and second-year growth of wetland sedges. New Phytologist, 166: 537-550. DOI: 10.1111/j.1469-8137.2005. 01320.x
- HAN W.X., FANG J.Y., REICH P.B., IAN WOODWARD F., WANG Z.H. 2011. Biogeography and variability of eleven mineral elements in plant leaves across gradients of climate, soil and plant functional type in China. Ecol. Lett., 14: 788-796.
- LAWNICZAK A.E. 2011. Response of two wetland graminoids to N:K supply ratios in a two-year growth experiment. J. Elem., 16(3): 421-436. DOI 10.5601/jelem.2011.16.3.07.
- OSTROWSKA A. 1987. Application of ANE value and shares of individual elements in this value for determining the difference between various plant species. In: Genetic aspects of plant mineral nutrition. 27-43, Dordrecht (Netherlands), Martinus Nijhoff Pub. (Plant and Soil). ISBN 90-2473494-0, p. 27-43.
- OSTROWSKA A., POREBSKA G. 2002. *Mineral composition of plants, interpretation and use in environmental protection*. Monograph. Institute of Environmental Protection, Warszawa, 1-165. (in Polish)
- OSTROWSKA A., POREBSKA G. 2009. Mineral composition of plants as an indicator of their quality in nitrogen stress condition. Ecol. Chem. Eng. A., 16: 983993.
- OSTROWSKA A., POREBSKA G. 2015. Assessment of the C/N ratio as an indicator of the decomposability of organic matter in forest soils. Ecol. Indic., 49: 104-109. http://dx.doi.org/10.1016/j. ecolind.2014.09.044
- POORTER H., BERGKOTTE M. 1992. Chemical composition of 24 wild species differing in relative growth rate. Plant Cell Environ., 15(2): 221-229.
- RIVAS-UBACH A., SARDANS J., PÉREZ-TRUJILLO M., ESTIARTE M., PENUELAS J. 2012. Strong relationship between elemental stoichiometry and metabolome in plants. Proc Natl Acad Sci USA, 109(11): 4181-4186. DOI: 10.1073/pnas.1116092109
- SARDANS J., RIAS-UBACH A., PENUELAS J. 2012. The elemental stoichiometry of aquatic and terrestrial ecosystems and its relationships with organismic lifestyle and ecosystem structure and function: a review and perspectives. Biogeochemistry, 111: 1-39.
- SIENKIEWICZ J., PORĘBSKA G., OSTROWSKA A., GOZDOWSKI D. Indicators of peat soil degradation in the Biebrza Valley, Poland. Wetlands. (in press)
- SIENKIEWICZ J., OSTROWSKA A., VOHLAND K., STRATMANN L., GRYGORUK M. 2014. Indicators for monitoring climate change-induced effects on habitats – a wetland perspective. In: Managing protected areas in Central and Eastern Europe under climate change. RANNOW S. and NEUBERT M. (eds). Adv. Global Change Res., 58: 77-94.

STERNER R.W., ELSER J.J. 2002. Ecological stoichiometry. Princeton University Press.

- STOLARSKI M., SZCZUKOWSKI S., TWORKOWSKI J. 2008. Biofuels from long standing energy crops. Energetyka, 1: 77-80. (in Polish)
- THOMAS S.C., MALCZEWSKI G. 2007. Wood carbon content of tree species in Eastern China: Interspecific variability and the importance of the volatile fraction. J. Environ. Manag., 85: 659-662.
- VAN DE WAAL D.B., SMITH V.H., DECLERCK S.A.J., STAM E.C.M., ELSER J.J. 2014. Stoichiometric regulation of phytoplankton toxins. Ecol. Lett., 17: 736-742.
- WANG M., MOORE T.R. 2014. Carbon, nitrogen, phosphorus, and potassium stoichiometry in an ombrotrophic peatland reflects plant functional type. Ecosystems, 17: 673-684. DOI: 10.1007/s10021-014-9752-x
- XIA J., WAN S. 2008. Global response patterns of terrestrial plant species to nitrogen addition. New Phytologist, 179: 428-439.